

Degradation and resilience in Louisiana salt marshes after the BP–Deepwater Horizon oil spill

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More than 2 y have passed since the BP–Deepwater Horizon oil spill in the Gulf of Mexico, yet we still have little understanding of its ecological impacts. Examining effects of this oil spill will generate much-needed insight into how shoreline habitats and the valuable ecological services they provide (e.g., shoreline protection) are affected by and recover from large-scale disturbance. Here we report on not only rapid salt-marsh recovery (high resilience) but also permanent marsh area loss after the BP–Deepwater Horizon oil spill. Field observations, experimental manipulations, and wave-propagation modeling reveal that (i) oil coverage was primarily concentrated on the seaward edge of marshes; (ii) there were thresholds of oil coverage that were associated with severity of salt-marsh damage, with heavy oiling leading to plant mortality; (iii) oil-driven plant death on the edges of these marshes more than doubled rates of shoreline erosion, further driving marsh platform loss that is likely to be permanent; and (iv) after 18 mo, marsh grasses have largely recovered into previously oiled, noneroded areas, and the elevated shoreline retreat rates observed at oiled sites have decreased to levels at reference marsh sites. This paper highlights that heavy oil coverage on the shorelines of Louisiana marshes, already experiencing elevated retreat because of intense human activities, induced a geomorphic feedback that amplified this erosion and thereby set limits to the recovery of otherwise resilient vegetation. It thus warns of the enhanced vulnerability of already degraded marshes to heavy oil coverage and provides a clear example of how multiple human-induced stressors can interact to hasten ecosystem decline.

geomorphology | multiple stressor | wetland | human impacts

Human activities severely threaten coastal ecosystems and the critical services they provide worldwide (1–4). Pollution from point-source release is often among the most intense of these anthropogenic stressors and can drive severe and rapid degradation of local habitats, such as seagrasses, mangroves, and coral reefs (e.g., refs. 5–8). Oil spills, in particular, pose a heightened threat to ecosystem health because they are unpredictable in space and time, and the resources needed to minimize impacts are often not immediately available (e.g., a containment cap for a well blowout) (9). Past oil spills in coastal habitats have led to immediate effects such as widespread animal die-offs and losses of ecosystem services (e.g., refs. 5, 7, 10–12) as well as longer-lasting effects, such as alteration of animal behaviors and persistence of oil-derived compounds in food webs (e.g., refs. 5–8, and 13–21).

In April 2010, well blowout on the seafloor below the BP-contracted Deepwater Horizon (BP-DWH) oil-drilling vessel, ~80 km off the Louisiana coast, led to the eventual release of an estimated 4,900,000 barrels of crude into Gulf of Mexico waters (22), some portion of which ultimately landed in nearby shoreline ecosystems (23). Various sources estimate that ~75 linear km of salt marsh in Louisiana experienced moderate to heavy oiling, the most of any state (24, 25) (Fig. S1). Gulf of Mexico coastal habitats are economically important, generating more than \$10 billion per year in revenues through fisheries and

tourism (4, 24, 26). Salt marshes, as one of the most common ecosystems in this region, are critical to maintaining these valuable ecosystem services (4, 26, 27). Past studies investigating effects of oil spills on salt marshes indicate that negative impacts on plants can be overcome by vegetation regrowth into disturbed areas once the oil has been degraded (8, 28–30). This finding suggests that marshes are intrinsically resilient to (i.e., able to recover from) oil-induced perturbation, especially in warmer climates such as the Gulf of Mexico, where oil degradation and plant growth rates may be high.

Here, we report on underappreciated indirect effects caused by the interaction between shoreline oiling and geomorphic feedbacks in salt marshes that dramatically reduce salt-marsh resilience to oil disturbance. In oil spill-impacted Louisiana marshes, we examined (i) the magnitude of oil contamination at different distances from the shoreline; (ii) the effects of this oil contamination on salt-marsh flora, fauna, and shoreline position; and (iii) the recovery of salt-marsh ecosystems after intense and localized oil coverage. To evaluate oil impacts on the marsh ecosystem, we used a multicomponent approach of (i) replicated, control-impact-paired time-series studies; (ii) shallow-water, oceanographic wave-propagation modeling; and (iii) in situ field experiments.

Results and Discussion

In October 2010, about 5 mo after initial oil contact, we surveyed marsh sites throughout Barataria Bay, LA, a coastal region that experienced some of the most extensive BP-DWH oil contamination (Fig. S1). Although interior marsh regions were intact (vegetation >15 m from the marsh edge), marsh shoreline habitats (<15 m from the marsh edge) were mixtures of apparently healthy and severely degraded, oil-impacted sites (i.e., muddy areas laden with oil-covered dead and horizontally laying, decaying grass stems). At oil-impacted sites ($n = 3$; Methods), we found abundant oil residues (up to 82% on an aerial basis; Figs. 1A and 2) on the marsh substrate, in contrast to the low levels of other known plant stressors (i.e., redox potential, soil salinities, fungal-farming snails; Figs. S2 and S3) that have driven previous marsh community die-offs (31–35). No oil residue was observed at our reference sites ($n = 3$) on either marsh plants or the substrate (Fig. 1A). In addition, levels of total polyaromatic hydrocarbons (PAHs; a proxy for oil residue abundance) found in the surface sediments at impacted sites was >100 times higher than concentrations found in reference marshes (Fig. 1B).

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in reference marshes were dead (Fig. 1*D*). Oil cover on marsh surfaces dropped precipitously at impacted sites at distances beyond 10 m from the shoreline, and live plant cover concomitantly increased to more than 50% (Fig. 1*A* and *C*). Levels of PAHs decreased beyond 15 m from the shoreline and were not statistically different from those at reference marsh sites (Fig. 1*B*). Above-ground plant and rhizome concentrations also increased beyond 15 m to match those found in reference marshes (Fig. 1*C* and *D*).

These data provide evidence of salt-marsh community die-off in the near-shore portion of the Louisiana shoreline after the BP-DWH oil spill because of high concentrations of oil at the edge of the marsh. Specifically, these findings suggest that the vegetation at the marsh edge, by reaching above the highest high-tide line in the microtidal environment of the Gulf of Mexico, blocked and confined incoming oil to the shoreline region of the marsh. This shoreline containment of the oil may have protected inland marsh but led to extensive mortality of marsh plants located from the marsh edge to 5–10 m inland and to sublethal plant impacts on plants 10–20 m from the shoreline, where plant oiling was less severe. This assertion is also supported by data from our field experiment and supplemental field surveys that assessed impacts of covering by oil (collected from the marsh surface) on the health of live marsh plants (Fig. 3, *Methods*, and *SI Methods*). Specifically, our studies revealed a nonlinear relationship between stem oil coverage and stem death and a threshold of oil coverage of ~65%, beyond which plant death occurred—the same high-oil coverage observed on plants and the marsh surface at the seaward edge of our impacted marsh sites and across an additional four oiled sites we surveyed (Fig. 3). Moreover, tests of alternative causes for marsh die-offs, such as drought, inundation, or grazers, all failed to explain the observed pattern of ecosystem loss (*SI Methods*). These data also suggest that the mechanism of the lethal effects of oil are more

likely derived from interference with respiration and photosynthesis (reviewed in ref. 8) than from direct toxicity because plant death only occurred at high levels of oil coverage.

Biogeomorphological Feedback. Oil concentrated on the marsh edge enhanced the rate of decline of Louisiana salt marshes, which are known to be degrading at an alarming pace (e.g., refs. 35–38). Specifically, erosion on the steep edges of these already receding marsh platforms was more than twice as high at oil-impacted sites than at reference sites during the period between October 2010 (~5 mo after oil was reported on Louisiana marshes) and October 2011 (Fig. 4*A*). After October 2011, erosion rates did not differ between impacted and reference sites (Fig. 4*A*). In nondegraded salt-marsh plant communities, belowground plant architecture is characterized by a complex network of underground roots and rhizomes generated by clonally reproducing plants (39–41). This elaborate root matrix helps to maintain shoreline structure and retard erosion by binding sediments and increasing concentrations of organic matter that act as adhesive agents (38, 42–45). Our results suggest that oil-generated death of this stabilizing root matrix at the edges of these marshes triggered a geomorphic response that led to accelerated erosion of the marsh edge, hastening the degradation of the elevated platform on which marsh vegetation depends (35, 46–50) (Fig. 4). Our study sites were all of similar physiographic character because there were no differences in the shallow-water slopes among reference and impacted sites ($P = 0.55$). We conducted numerical simulations of wind-generated wave growth and propagation with SWAN (51) to ensure that differences in observed erosion rates between impacted and reference sites were not attributable to a predisposition of the impacted sites to higher erosion rates (Fig. S4). In fact, the model results demonstrate that our reference sites receive slightly higher wave-energy fluxes than the impacted sites (Fig. S4).

Ecosystem Resilience and Degredation. Despite the deleterious effects of the oil spill on marsh vegetation and erosion rates, we found clear evidence for recovery processes. In our transect surveys at impacted sites in April 2011 (~11 mo after oil coverage occurred on these marshes), we documented significant increases in plant cover (up to 33%, on average ~20%; Fig. 4*C–F*) in areas that were barren and had nearly complete die-back of roots during our initial surveys (Figs. 1*C* and 2). Because we observed no seedling establishment in impacted areas, this recovery likely occurred via plant lateral regrowth (i.e., clonal growth) originating from interior marsh areas where plants were less affected or from nearby, small remnant patches in the impacted areas (52). This clonal regrowth of marsh plants continued throughout the summer of 2011, with full recovery of the marsh plant cover occurring sometime between October 2011 and January 2012, ~1.5 y after the oil spill (Fig. 4). As predicted from past studies (36, 42–45), plant shoreline reestablishment suppressed the observed accelerated erosion rates at impacted sites to values not significantly different from those at reference sites (Fig. 4 and Fig. S4). However, no plant recovery was observed in the marsh platforms lost to accelerated erosion, and marsh plants that were transplanted into these eroded areas in June 2011 died within 2 mo (Fig. S5 and *SI Methods*), whereas those planted in nonoiled areas of both reference and impacted sites remained alive as of January 2012. Our observations agree with past modeling studies revealing that accelerated erosion on marsh cliffs in Louisiana reduces substrate surface to subtidal elevations and thereby prohibits the recovery of salt-marsh vegetation (48, 53, 54). However, our observations are in contrast to a number of previous studies from other regions that found, under natural conditions, salt-marsh vegetation is resilient to this stress and reestablishes seaward of eroding edges (33, 48, 55). The absence of recovery of marsh vegetation seaward of the

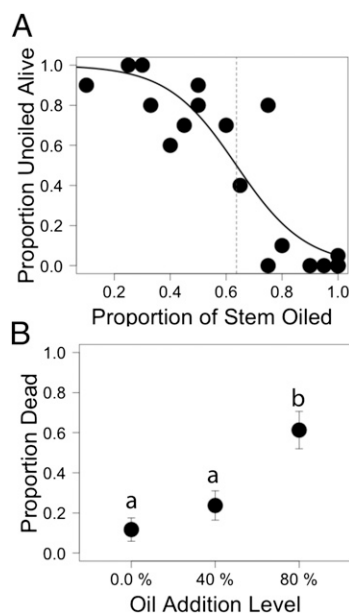


Fig. 3. Oil cover versus plant death as assessed from field observations and from manipulations. (*A*) Field observations of level of oil on individual plants and resultant plant death, indicated by blade browning. The proportion of plant stems that were green and alive is greater than the proportion dead (i.e., indicating improved health) when oil coverage dropped below 64%. (*B*) Observation of blade browning 30 d after a treatment of oil coverage of 40% or 80% of the plant's height ($n = 6$ per oil-addition treatment). Superscript letters indicate treatments that were significantly different ($P < 0.05$) based on Tukey's HSD post hoc comparison. Error bars are SEs.

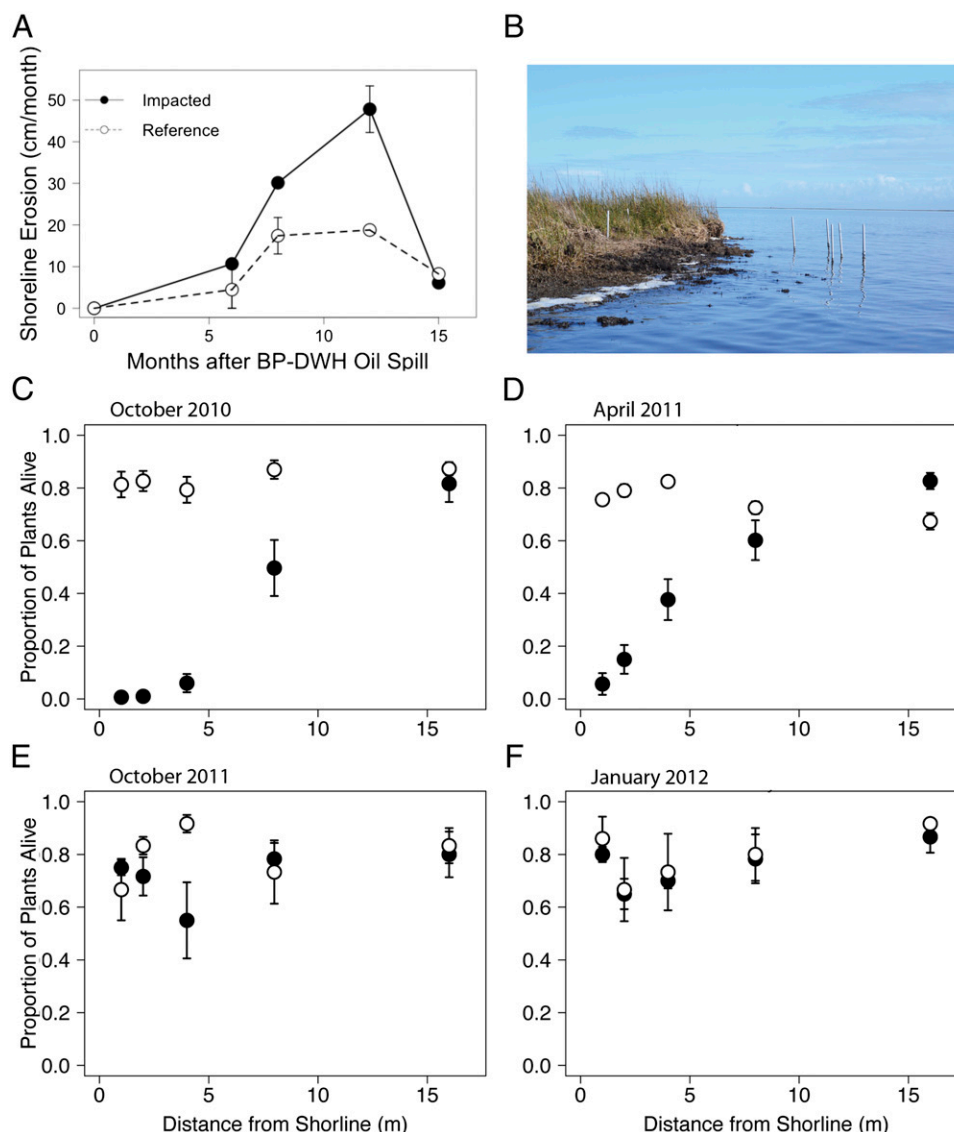


Fig. 4. (A) There was a significant increase in average lateral shoreline erosion rate between reference and impacted sites ($P = 0.007$) based on measurements at each site type. Error bars are SEs, and unseen error bars are smaller than symbols. (B) Photo of erosion monitoring poles at an impacted site. Right-most PVC poles were installed to mark the marsh platform edge, and retreat of the marsh from this initial starting point is apparent. (C–F) Comparison between average percentage of plants alive at four times at impacted and reference sites from 0 to 15 m from the shoreline ($n = 3$). There was a significant effect of the presence of oil, the distance from shoreline, and time ($P < 0.0001$), with much lower plant coverage near shore for impacted sites than for reference sites during October 2010 and April 2011 but with similar levels of coverage near shore during October 2011 and January 2012 at these sites. Plant coverage was similar for all sites and times at greater distances from the shoreline beyond 10 m from the marsh edge.

retreating marsh cliffs in Barataria Bay underscores that this oil spill has decreased the resilience of these marshes by triggering accelerated substrate erosion, which, in turn, reduces the overall area that can be recolonized by plants (Fig. S5).

Our results suggest that there are reasons for both optimism and concern about the impact of this oil spill on Mississippi deltaic marshes of Louisiana. On one hand, our results reveal that marsh vegetation displays remarkable resilience to oil spills by concentrating and confining the effects of oil to the marsh edge, recovering fully in noneroded areas after ~ 1.5 y, and suppressing, through this recolonization, further accelerated erosion rates along the shoreline. The lack of oil on the marsh surface or on grasses at distances greater than 15 m from the shoreline at any site (Fig. 1A) suggests that incoming oil sheens were contained and prevented from moving into interior marshes by a baffling wall of live and dying salt-marsh grasses, a process

that in itself increases the resistance of the extensive marsh ecosystem to oil spill. However, this resistance comes at a high cost for the impacted areas because marsh grass die-off and subsequent sediment exposure to waves resulted in a more than doubling of the rate of erosion of the intertidal platform, leading to permanent marsh ecosystem loss (Fig. 4). Specifically, we observed an average rate of marsh shoreline retreat of ~ 1.38 $\text{m}\cdot\text{y}^{-1}$ (Fig. 4A) at our reference sites, a level consistent with that reported in other studies for this area of Louisiana ($0.8\text{--}1.3$ $\text{m}\cdot\text{y}^{-1}$) (38). This already high rate of shoreline retreat increased by more than 125% to ~ 3.0 $\text{m}\cdot\text{y}^{-1}$ at oil-impacted sites. Indeed, the extent of habitat loss could have been even more severe if a large storm or hurricane had coincided with the period of increased shoreline exposure after oil-driven die-off. In fact, the rapid shoreline retreat rates observed between April and October 2011 may be attributable to the effects of Hurricane Lee in September

2011, which made landfall immediately west of Barataria Bay, bringing high winds and surge-related flooding to the region. The highly elevated erosion rates after oil-driven marsh grass die-off observed in this study (Barataria Bay and Bay Jimmy; Fig. S1) are likely general and can be extrapolated to the other marshes in Louisiana that also experienced moderate-to-heavy oil coverage because these marshes are also typically characterized by erosive edges (33, 55–57).

More broadly, our results reveal that multiple stressors are interacting in Louisiana marshes to hasten ecosystem decline. Louisiana experiences some of the highest rates of salt-marsh loss in North America (~75 km²/y) as a result of natural subsidence and channelization of the Mississippi River, which reduces sediment supply to the coast, causing submersion of the marsh interiors and formation of erosion-prone, cliffed edges (31, 37, 54). Our observations and experimental work demonstrate that intense oil coverage of these already degraded marsh edges interacts with preexisting sediment-limitation stress to amplify permanent habitat loss along the marsh margins. The edges of healthy marshes are typically characterized by more gently sloping banks and therefore tend to be more resistant to erosion than subsiding, deltaic-plain marshes in Louisiana that are often characterized by erosive cliff edges (33, 55–57). This study highlights the enhanced vulnerability of these already degraded marshes to heavy oil coverage associated with oil spills and provides a clear example of how multiple human-induced stressors can interact to hasten the loss of a critical marine ecosystem and the services it provides.

Although the amount of increased erosion caused by loss of marsh plants from oiling is, in many ways, specific to these oil-impacted areas in Louisiana (e.g., because of the microtidal environment and long-term sediment deprivation from river channelization), the mechanisms underlying these results can likely be extended to other oil-impacted coastal salt marshes that are also characterized by erosive edges or cliffs. In fact, erosion of marshes in response to oil coverage has been observed or suggested in studies in New England, Florida, and Louisiana (58–62). Our study goes one step further to show that direct

vegetation die-off is the primary result of heavy oil coverage, but that erosion caused by biogeomorphological feedback and subsequent habitat loss may ultimately determine the long-term effect of oil pollution in salt marshes. Future research should focus on how the interplay between biological and geomorphological processes affects the vulnerability of salt marshes facing multiple anthropogenic stressors.

Methods

All work was conducted in *Spartina alterniflora* (cordgrass)-dominated salt marshes in Barataria Bay, LA, one of the heaviest-impacted areas after the BP-DWH oil spill (Fig. S1). Sites were identified as either “impacted” or “reference.” Impacted sites had substrate that were denuded and/or laden with dead and decaying cordgrass stems. Reference sites were dominated by standing live cordgrass plants. All impacted sites were located in the northeast corner of Barataria Bay, near Bay Jimmy, which received large amounts of oil coverage because of prevailing winds and currents after the BP-DWH oil spill (Table S1). Two reference sites were selected in the northwest corner of the bay, east of Hackberry Bay, and an additional reference site was located in Grand Isle State Park on the south side of the bay. All sites were located within 30 km of each other. All but the Grand Isle State Park site faced outward into the same portion of open Barataria Bay, thus experiencing similar tidal fluctuations and weather conditions. Because the Grand Isle site was relatively protected from wave action, this site was not included in our comparison of erosion rates between reference and impacted sites (see Table S1 for coordinates of reference and impacted sites). At both reference and impacted sites, we ran an oil-addition and a marsh plant transplant experiment, and we conducted surveys in which oil coverage, PAH concentrations, live plant (*S. alterniflora*) coverage and survival, invertebrate abundances, pH, redox and salinity, as well as erosion potential and erosion rates were quantified (for detailed sampling and analytical methods, see *SI Methods*).

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- Griffis RB, Kimball KW (1996) Ecosystem approaches to coastal and ocean stewardship. *Ecol Appl* 6(3):708–712.
- Scavia D, et al. (2002) Climate change impacts on U.S. coastal and marine ecosystems. *Estuaries Coasts* 25(2):149–164.
- Waycott M, et al. (2009) Accelerating loss of seagrasses across the globe threatens coastal ecosystems. *Proc Natl Acad Sci USA* 106(30):12377–12381.
- Mitsch WJ, Gosselink JG (2000) The value of wetlands: Importance of scale and landscape setting. *Ecol Econ* 35(1):25–33.
- Cubit JD, et al. (1987) An oil spill affecting coral reefs and mangroves on the Caribbean coast of Panama. *Proceedings of the 1987 Oil Spill Conference* (American Petroleum Institute, Washington, DC), pp 401–406.
- Duke NC, Pinzón M, Zuleika S, Prada T, Martha C (1997) Large scale damage to mangrove forests following two large oil spills in Panama. *Biotropica* 29:2–14.
- Jackson JBC, et al. (1989) Ecological effects of a major oil spill on Panamanian coastal marine communities. *Science* 243(4887):37–44.
- Pezeshki SR, Hester MW, Lin Q, Nyman JA (2000) The effects of oil spill and clean-up on dominant US Gulf coast marsh macrophytes: A review. *Environ Pollut* 108(2):129–139.
- Lin Q, Mendelsohn IA (1998) The combined effects of phytoremediation and biostimulation in enhancing habitat restoration and oil degradation of petroleum contaminated wetlands. *Ecol Eng* 10(3):263–274.
- Garrott RA, Eberhardt LL, Burn DM (1993) Mortality of sea otters in Prince William Sound following the Exxon Valdez oil-spill. *Mar Mamm Sci* 9(4):343–359.
- Jewett SC, Dean TA, Smith RO, Blanchard A (1999) “Exxon Valdez” oil spill: Impacts and recovery in the soft-bottom benthic community in and adjacent to eelgrass beds. *Mar Ecol Prog Ser* 185:59–83.
- Jackson JBC, et al. (2001) Historical overfishing and the recent collapse of coastal ecosystems. *Science* 293(5530):629–637.
- Andrade ML, Covelio EF, Vega FA, Marcet P (2004) Effect of the Prestige oil spill on salt marsh soils on the coast of Galicia (northwestern Spain). *J Environ Qual* 33(6):2103–2110.
- Day RH, et al. (1997) Effects of the Exxon Valdez oil spill on habitat use by birds in Prince William Sound, Alaska. *Ecol Appl* 7(2):593–613.
- Monson DH, Doak DF, Ballachey BE, Johnson A, Bodkin JL (2000) Long-term impacts of the Exxon Valdez oil spill on sea otters, assessed through age-dependent mortality patterns. *Proc Natl Acad Sci USA* 97(12):6562–6567.
- Murphy ML, Heintz RA, Short JW, Larsen ML, Rice SD (1999) Recovery of pink salmon spawning areas after the Exxon Valdez oil spill. *Trans Am Fish Soc* 128(5):909–918.
- Ocon CS, Rodrigues Capitulo A, Paggi AC (2008) Evaluation of zoobenthic assemblages and recovery following petroleum spill in a coastal area of Rio de la Plata estuarine system, South America. *Environ Pollut* 156(1):82–89.
- Peterson CH, et al. (2003) Long-term ecosystem response to the Exxon Valdez oil spill. *Science* 302(5653):2082–2086.
- Short JW, et al. (2004) Estimate of oil persisting on the beaches of Prince William Sound 12 years after the Exxon Valdez oil spill. *Environ Sci Technol* 38(1):19–25.
- Trust KA, Esler D, Woodin BR, Stegeman JJ (2000) Cytochrome P450 1A induction in sea ducks inhabiting nearshore areas of Prince William Sound, Alaska. *Mar Pollut Bull* 40(5):397–403.
- Culbertson JB, et al. (2007) Long-term biological effects of petroleum residues on fiddler crabs in salt marshes. *Mar Pollut Bull* 54(7):955–962.
- McNutt M, et al. (2011) Assessment of Flow Rate Estimates for the Deepwater Horizon/Macondo Well Oil Spill. Flow Rate Technical Group Report to the National Incident Command, Interagency Solutions Group. Available at <http://www.doi.gov/deepwaterhorizon/loader.cfm?csModule=security/getfile&PageID=237763>. Accessed May 4, 2012.
- Thibodeaux LJ, et al. (2011) Marine oil fate: Knowledge gaps, basic research, and development needs; A perspective based on the Deepwater Horizon spill. *Environ Eng Sci* 28(2):87–93.
- Oil Spill Commission (2011) Deep Water: The Gulf Oil Disaster and the Future of Offshore Drilling. Available at <http://www.gpo.gov/fdsys/pkg/GPO-OILCOMMISSION/content-detail.html>. Accessed March 2011.
- Lehner P, Deans B (2010) *In Deep Water: The Anatomy of a Disaster, the Fate of the Gulf, and Ending Our Oil Addiction* (Experiment LLC, New York).
- Engle VD (2011) Estimating the provision of ecosystem services by Gulf of Mexico coastal wetlands. *Wetlands* 31(1):179–193.
- Barbier EB, Heal GM (2006) Valuing ecosystem services. *Economists’ Voice* 3(3):2.

28. Alexander SK, Webb JW (1985) Seasonal response of *Spartina alterniflora* to oil. *Proceedings of the 1985 Oil Spill Conference (Prevention, Behavior, Control, Cleanup)* (American Petroleum Institute, Washington, DC), pp 355–357.
29. DeLaune RD, Pezeshki SR, Jugsujinda A, Lindau CW (2003) Sensitivity of US Gulf of Mexico coastal marsh vegetation to crude oil: Comparison of greenhouse and field responses. *Aquat Ecol* 37(4):351–360.
30. Lytle JS (1975) Fate and effect of crude oil on an estuarine pond. *Proceedings of the Joint Conference on Prevention and Control of Oil Pollution* (American Petroleum Institute, Washington, DC), pp 595–600.
31. Gedan KB, Silliman BR (2009) Patterns of salt marsh loss within coastal regions of North America: Presettlement to present. *Human Impacts on Salt Marshes: A Global Perspective*, eds Silliman BR, Grosholz ED, Bertness MD (Univ of California Press, Berkeley), pp 253–266.
32. Kennish MJ (2002) Environmental threats and environmental future of estuaries. *Environ Conserv* 29(1):78–107.
33. Kirwan ML, et al. (2010) Limits on the adaptability of coastal marshes to rising sea level. *Geophys Res Lett* 37(23):L23401.
34. Silliman BR, van de Koppel J, Bertness MD, Stanton LE, Mendelsohn IA (2005) Drought, snails, and large-scale die-off of southern U.S. salt marshes. *Science* 310(5755):1803–1806.
35. DeLaune RD, Nyman JA, Patrick WH, Jr. (1994) Peat collapse, ponding and wetland loss in a rapidly submerging coastal marsh. *J Coast Res* 10:1021–1030.
36. Hatton RS, DeLaune RD, Patrick WH, Jr. (1983) Sedimentation, accretion, and subsidence in marshes of Barataria Basin, Louisiana. *Limnol Oceanogr* 28(3):494–502.
37. Reed DJ (1989) Patterns of sediment deposition in subsiding coastal salt marshes, Terrebonne Bay, Louisiana: The role of winter storms. *Estuaries Coasts* 12(4):222–227.
38. Wilson CA, Allison MA (2008) An equilibrium profile model for retreating marsh shorelines in southeast Louisiana. *Estuar Coast Shelf Sci* 80(4):483–494.
39. Gallagher JL, Plumley FG (1979) Underground biomass profiles and productivity in Atlantic coastal marshes. *Am J Bot* 66(2):156–161.
40. Smith KK, Good RE, Good NF (1979) Production dynamics for above and belowground components of a New Jersey *Spartina alterniflora* tidal marsh. *Estuar Coast Mar Sci* 9(2):189–201.
41. Valiela I, Teal JM, Persson NY (1976) Production and dynamics of experimentally enriched salt marsh vegetation: Belowground biomass. *Limnol Oceanogr* 21(2):245–252.
42. Knutson PL, Ford JC, Inskeep MR, Oyler J (1981) National survey of planted salt marshes (vegetative stabilization and wave stress). *Wetlands* 1(1):129–157.
43. Woodhouse WW, Jr., Seneca ED, Broome SW (1974) *Propagation of Spartina alterniflora for Substrate Stabilization and Salt Marsh Development* (U.S. Army Corps of Engineers Coastal Engineering Research Center, Fort Belvoir, VA), Tech Memo 46.
44. Woodhouse WW, Jr., Seneca ED, Broome SW (1976) *North Carolina State Univ at Raleigh. Propagation and Use of Spartina alterniflora for Shoreline Erosion Abatement* (U.S. Army Corps of Engineers Coastal Engineering Research Center, Fort Belvoir, VA), Tech Rep 76-2.
45. Stagg CL, Mendelsohn IA (2011) Controls on resilience and stability in a sediment-subsided salt marsh. *Ecol Appl* 21(5):1731–1744.
46. Cahoon DR, et al. (2003) Mass tree mortality leads to mangrove peat collapse at Bay Islands, Honduras after Hurricane Mitch. *J Ecol* 91(6):1093–1105.
47. Hughes ZJ, et al. (2009) Rapid headward erosion of marsh creeks in response to relative sea level rise. *Geophys Res Lett* 36:1–5.
48. Kirwan ML, Murray AB (2007) A coupled geomorphic and ecological model of tidal marsh evolution. *Proc Natl Acad Sci USA* 104(15):6118–6122.
49. Marani M, D'Alpaos A, Lanzoni S, Carniello L, Rinaldo A (2007) Biologically-controlled multiple equilibria of tidal landforms and the fate of the Venice lagoon. *Geophys Res Lett* 34:1–5.
50. Smith SM (2009) Multi-decadal changes in salt Marshes of Cape Cod, MA: Photographic analyses of vegetation loss, species shifts, and geomorphic change. *Northeast Nat* 16(2):183–208.
51. Booij N, Ris R, Holthuijsen L (1999) A third-generation wave model for coastal regions. 1. Model description and validation. *J Geophys Res* 104(C4):7649–7666.
52. Angelini C, Silliman BR (2012) Patch size-dependent community recovery after massive disturbance. *Ecology* 93:101–110.
53. Mendelsohn IA, Morris JT (2002) Eco-physiological controls on the productivity of *Spartina alterniflora*. *Concepts and Controversies in Tidal Marsh Ecology*, eds Weinstein NP, Kreeger DA (Kluwer Academic, Boston), pp 59–80.
54. Blum MD, Roberts HH (2009) Drowning of the Mississippi Delta due to insufficient sediment supply and global sea-level rise. *Nat Geosci* 2(7):488–491.
55. Mariotti G, Fagherazzi S (2010) A numerical model for the coupled long-term evolution of salt marshes and tidal flats. *J Geophys Res* 115:F01004.
56. Kirwan ML, Murray AB, Donnelly JP, Corbett DR (2011) Rapid wetland expansion during European settlement and its implication for marsh survival under modern sediment delivery rates. *Geology* 39(5):507–510.
57. Möller I, Spencer T (2002) Wave dissipation over macro-tidal saltmarshes: Effects of marsh edge typology and vegetation change. *J Coast Res* 36:506–521.
58. DeLaune RD, White JR (2011) Will coastal wetlands continue to sequester carbon in response to an increase in global sea level?: A case study of the rapidly subsiding Mississippi river deltaic plain. *Clim Change* 110(1-2):1–18.
59. Hampson GR, Moul ET (1978) No. 2 fuel oil spill in Bourne, Massachusetts: Immediate assessment of the effects on marine invertebrates and a 3-year study of growth and recovery of a salt marsh. *J Fish Res Board Can* 35(5):731–744.
60. Hartig EK, Gornitz V, Kolker A, Mushacke F, Fallon D (2002) Anthropogenic and climate-change impacts on salt marshes of Jamaica Bay, New York City. *Wetlands* 22(1):71–89.
61. McBride RA, Byrnes MR (1997) Regional variations in shore response along barrier island systems of the Mississippi River delta plain: Historical change and future prediction. *J Coast Res* 13:628–655.
62. Peacock EE, et al. (2007) The 1974 spill of the Bouchard 65 oil barge: Petroleum hydrocarbons persist in Winsor Cove salt marsh sediments. *Mar Pollut Bull* 54(2):214–225.